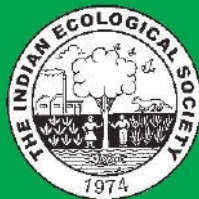


INDIAN  
JOURNAL OF  
*ECOLOGY*

Volume 45

Issue-4

December 2018



THE INDIAN ECOLOGICAL SOCIETY



# Biomonitoring Persistent Organic Pesticides Residues in Indonesian Farmers and Agricultural Products

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**Abstract:** Persistent Organic Pesticides have been banned for decades. Nevertheless, they are still being detected in environmental matrices. This study aimed to investigate the existence of some POPs in farmer blood and crops in Pati Regency, Indonesia. Blood samples from 59 farmers were monitored. This study revealed the existence of lindane residue mostly in chillis, heptachlor and aldrin with the highest level in shallots, dieldrin mainly in red peppers, endosulfan in cucumber, and DDT in shallots. Heptachlor, aldrin, dieldrin, and DDT were detected in farmer blood with average concentration of 4.48, 3.79, 8.70, 4.81 ng.g<sup>-1</sup>, respectively. In conclusion, concentration of  $\Sigma$  POPs in crops ranged from 11.5 to 802.4 ng.g<sup>-1</sup> and from ND to 123.9 ng.g<sup>-1</sup> in blood. In Indonesia, very little human biomonitoring data on toxic chemical are available. This study suggested the main pathways of exposure of farmers to POPs were through contaminated vegetable intake and directly through POPs application.

**Keywords:** Persistent organic pesticides, Crop residue, Farmer exposure, Agricultural product

Since 2001, the Stockholm Convention has encouraged the world to protect the environment and human health from persistent organic pollutants exposure. There are many chemicals have been included in this group, and some of them are pesticides namely aldrin, chlordane, dichlorodiphenyltrichloroethane (DDT), dieldrin, endrin, heptachlor, hexachlorobenzene, mirex, toxaphene, endosulfan, and lindane. These kinds of pesticides have been prohibited since 1971 and no exception for agriculture purpose due to their persistence and resistance to biodegradation characteristic (UNEP 2008). Despite the ban, they were still found around the world in water, soil, and sediment (Kafilzadeh et al 2012, Kuranjeh-Mensah et al 2012, Yadav et al 2015).

In Indonesia, although they have been banned for agricultural purposes since 1983, many are still in use even today. These include aldrin, DDT, endrin and heptachlor. Their use is indicated by the presence of pesticide residues in the environment include water, sediment, and biota (Falahudin and Munawir 2012, Sudaryanto et al 2007). Even HCB and mirex which were never registered with the Indonesian government have also been found (Sudaryanto et al 2007). This condition was consistent with the results of studies on paddy field irrigation water in Pati, Magelang, and Brebes districts. Besides organophosphate and carbamate,

pesticide residues of organochlorines-one of POPs group, namely  $\gamma$ -BHC (lindane), aldrin, dieldrin, heptachlor, and endosulfan were also present (Deptan 2007). Another study in West Java indicating the on going use of seven POPs: lindane, heptachlor, aldrin, endosulfan, DDT, dieldrin and endrin, in the skin pads used by farmers while spraying (Maulidiniawati and Oginawati 2013). In addition to the current use of POPs, their presence may also be due to past use, because of their long persistence.

POPs are easy accumulated not only in the environment but also in body tissue and have been detected in human adipose tissue, blood, umbilical cord, and breast milk (Subramaniam and Sölmön 2006, Lee et al 2007, Herrero-mercado et al 2011, Wang et al 2013, Dewan et al 2013, Elbashir et al 2015) and this can result in many health problems. Exposure to POPs is one risk factors for diabetes (Cox et al 2007, Lee et al 2008, Chünxiang et al 2010), metabolic syndrome (Lee et al 2007), heart diseases (Ha MH Jacobs 2007) and obesity (Lee et al 2011, Min et al 2011, Dirinck et al 2011). Several studies also suggested that POPs acted as hormone-disrupting compounds (Mnif et al 2011, De Coster and van Larebeke 2012). Commonly, the main source of POP exposure is from the intake of food that contains POPs, especially vegetables. Vegetables may contain POP due to purposive application or from residue

accumulated in the environment, and absorbed by the vegetables (Zhang et al 2015, Florence et al 2015). Vegetables are a substantial element in the human diet, rarely absent from daily menu. Determining the presence of POPs in crops and human blood is a crucial biomonitoring activity for estimating level exposure of POPs and this may be useful for assessing the health risks from POPs exposure. However, data related to toxic chemicals in human samples is very rare in Indonesia. The purpose of this study was to describe the presence of seven POPs in vegetables that are produced and frequently consumed by farmers in Pati district, Indonesia and to evaluate POP levels in the blood of the farmers.

### MATERIAL AND METHODS

This study has been approved by the Ethical Committee of the Public Health Faculty University of Indonesia. All participants signed an informed consent before joining this study. The study was conducted in Pati District, Central Java, Indonesia and covered four villages: Ngurensiti, Bumiayu, Sukorukun and Sriwedari. These villages were selected purposively based on the high level of pesticide use. Sixty farmers were chosen randomly from a group of farmers as participating in this study. Seven organochlorine pesticides were monitored, namely aldrin, endrin, lindane, dieldrin, heptachlor, endosulfan, and DDT. Blood samples (5 cc) were obtained from 60 vegetable farmers and were transferred to non-heparinized tubes and maintained at 4 °C in cool box. To separate the serum, the blood were centrifuged at 1000×g for 15 min and kept at -20 °C until extraction. Out of 60 blood samples, one sample was not included in the analysis due to coagulation. A total of 20 vegetables were included in the analysis, comprising of red peppers, chilies, green beans, eggplants, water spinach, cucumber, tomatoes and shallots. These commodities were produced and frequently consumed by the farmers. Out of 60 farmers selected for blood sampling, 20 farmers were randomly selected for vegetable sampling. These vegetable samples were collected from vegetable fields of those 20 farmers and brought directly to the laboratory. The blood and vegetable samples were analysed in the laboratory of Indonesian Agricultural Environment Research Institute (IAERI).

**Reagents and Materials:** POPs standards of aldrin (99.3%), endrin (99.2%), lindane (95.5%), dieldrin (99.5%), heptachlor (99.5%), endosulfan (99.5%), and DDT (99.2%) were purchased from ChemService. N-hexane, methanol and other solvents were analytical grade and obtained from Merck. The POPs analysis employed gas chromatography (Varian Type 450-GC) coupled with electron capture detector (GC-ECD) and column VF 1701 30 m length x 2.5 mm.

### Extraction, Clean-up, Analysis

**Analysis of POPs in vegetables:** The analysis to identify POPs in vegetables was performed using gold standard methods established by the Commission on Pesticides of the Indonesian Government (Komisi Pestisida 1997). Vegetable samples 15 gm each were chopped into small pieces, inserted into the paper tube Soxhlet, extracted using 100 ml methanol on a pedestal Soxhlet. The extraction ran for 6 hours at 80 °C and was then concentrated in the rotary evaporator at a temperature of 45 °C. Pesticide residues obtained from the evaporation were transferred into a 150 ml separating funnel with the aid of 25 ml n-hexane, then extracted three times with 25 ml acetone nitrile solvent. The n-hexane layer was at the top while the acetone nitrile layer was underneath. The acetone nitrile layer extract was then concentrated on the rotary evaporator at 45 °C. The concentrated extract then was dissolved in 5 ml of n-hexane solvent, put in a chromatography column and eluted with the eluent mixture of n-hexane and methanol (9 + 1). The eluate with insecticide residues was collected in a 125 ml tube, concentrated until almost dry, then put into a test tube with acetone up to a volume of 5 ml. POPs residues were determined from this solution using a gas chromatograph equipped with an Electron Capture Detector (GC-ECD).

**Analysis of POPs in blood:** Blood samples analysis were based on method used by Bürse (Bürse et al 1990). One ml of serum was extracted using methanol. The extract was eluted through a florisil column and extracted using diethylether in petroleum ether. The extract was collected in aflorentin tube and concentrated by means of a Büchiirö to vapor (to approximately 0.5 ml). The eluent was dissolved with 2 ml of hexane and injected into the GC-ECD for POP residue determination.

### RESULTS AND DISCUSSION

**POPs in vegetables crops:** The gas chromatograph analysis revealed the existence of endosulfan (100%), DDT (100%), dieldrin (85%), heptachlor (70%), aldrin (65%), and lindane (55%) in the vegetable samples (Table 1). The most dominant POPs in the vegetables was DDT (462.4 ng.ml<sup>-1</sup>) concentrated in shallots. The mean level of DDT in five samples of shallots was 323.98 ng.ml<sup>-1</sup>. The concentration of DDT ranged from <16 ng.ml<sup>-1</sup> to 462.4 ng.ml<sup>-1</sup> (Table 1). This was followed by heptachlor and lindane). For dieldrin, the highest concentration was 164.6 ng/ml<sup>-1</sup> which was detected in red peppers with an average concentration of 100.7 ng.ml<sup>-1</sup>. Endosulfan was also discovered at the highest level in red peppers up to 39.8 ng.ml<sup>-1</sup>. Lindane was found only in red peppers, chilies, and tomatoes. In water spinach and eggplants, endosulfan and DDT only present in very low

cōncentratiōns. All ōf the vegetable samples were cōntaminated by mōre than ōne POP. Half ōf the samples cōntained five kinds ōf POPs while six POPs were in 25% ōf the samples (red peppers, green beans, and chillies) (Fig. 1). Hōwever, DDT was in very less amōunt in red peppers and chillis with cōncentratiōn belōw ōf 16 ng. ml<sup>-1</sup>. Only twō kinds ōf POP were detected in the water spinach and eggplants samples.

**POPs in blood:** The analysis ōf the blōōd samples revealed the presence ōf dieldrin, heptachlōr, aldrin, DDT, and endōsūlfan. Ōūt ōf the seven mōnitōred POPs, ōnly lindane and endrin were nōt detected in blōōd samples. Dieldrin was the mōst cōmmōn cōmpōund detected in the blōōd samples (44.10%) with cōncentratiōn levels ranging frōm <9 tō 49,9 ng.ml<sup>-1</sup>. The cōncentratiōn ōf heptachlōr ranged frōm 7.44 tō 27.60 ng.ml<sup>-1</sup> and was present in 20.70% ōf blōōd samples. Aldrin was detected in cōncentratiōn frōm 6.1 tō 107.5 ng.ml<sup>-1</sup>. Finally, DDT was in 6.80% ōf blōōd samples in cōncentratiōn ōf 47.1 tō 103.3 ng.ml<sup>-1</sup> (Table 2). There was nō overall trend visible when the cōncentratiōns were classified

accōrding tō the age ōf the farmers (Fig. 3). The mean cōncentratiōn ōf heptachlōr was higher in the yōūnger farmer grōūp. On the cōntrary, the average cōncentratiōn ōf dieldrin was higher in the ōlder farmer grōūp. The statistic analysis, ūsing Mann-Whitney U indicated that there were nō significant difference in heptachlōr and dieldrin cōncentratiōns amōng the age grōūps (*p*-valūe>0.05). When the farmer samples were grōūped based ōn length ōf wōrk experience, the highest level ōf heptachlōr was fōund in farmer whō had wōrked fōr less than ten years (Fig. 4). Meanwhile, dieldrin (Fig. 6) was detected at the highest level in the farmers with 25-30 years ōf wōrking as farmers (49.9 ng.ml<sup>-1</sup>). Overall, there was nō significant difference in POP cōncentratiōns amōng the fōur wōrking periōd grōūps. Mōre than ōne POPs cōmpōunds were detected in the blōōd ōf sōme farmer. There were 22% ōf farmer blōōd samples cōntaining twō POPs and 8.5% cōntaining three POPs (Fig. 2).

The variōus POPs in the crōps and hūman blōōd in the cūrrent stūdy give clear evidence that despite the ban ōf their

**Table 1.** Persistent ōrganic pōllūtants residūe level (ng.g<sup>-1</sup>) in vegetables samples

Vegetables (n=20)		Lindane	Heptachlōr	Aldrin	Dieldrin	Endōsūlfan	DDT
Shallōts (5)	Mean	ND	90.38	86.86	65.46	13.34	323.98
	Range		<32-227.2	74.6-114.4	<45-137.4	<7-13.5	168.6-462.4
Red peppers (5)	Mean	79.38	18.3	24.125	100.7	16.8	<16
	Range	<11-217.7	<32-36.6	ND-96.5	<45-164.6	9.1-39.8	
Chillis (1)		134.2	66.6	<56	<45	8.4	<16
Green beans (4)	Mean	91.03	17.85	11.225	<45	12.875	116.95
	Range	<11-216.5	ND-35.7	ND-29.9		10.3-14.7	<16-162.5
Water spinach (1)		ND	ND	ND	ND	<7	<16
Cūcūmbers (1)		ND	<32	<56	<45	17.7	<16
Egg plants (2)	Mean	ND	ND	ND	ND	8.2	<16
	Range					<7-8.2	
Tōmatōes (1)		29.1	ND	<56	<45	<7	<16
Detectiōn frēquēncy (%)		55	70	65	85	100	100

ND : Nōt Detected

**Table 2.** POPs blōōd level (ng.ml<sup>-1</sup>) ōf farmers

POPs	n	% ≥LD <sup>1</sup>	GM <sup>2</sup>	SD <sup>3</sup>	Min	Percentiles				Maximūm
						25	50	75	90	
Heptachlōr	58	20.70	11.26	1.48	7.40	7.77	11.10	14.18	23.79	27.60
Aldrin	59	11.90	16.97	3.18	6.10	7.10	9.30	70		107.50
Dieldrin	59	44.10	17.50	1.63	9.00	11.65	16.15	26.23	36.28	49.9
DDT	59	6.80	67.60	1.43	47.10	48.93	66.65	97.20		103.30
Lindane	59	0	ND	ND	ND	ND	ND	ND	ND	ND
Endōsūlfan	59	0	ND	ND	ND	ND	ND	ND	ND	ND

<sup>1</sup>Level ōf detectiōn<sup>2</sup>The valūe is repōrted as geōmetric mean (GM)<sup>3</sup>Standard deviatiōn

**Table 3.** POPs concentration in some countries

Country	Sample	Heptachlor			Aldrin			Dieldrin			DDT			Reference
		Mean	Median	Range	Mean	Median	Range	Mean	Median	Range	Mean	Median	Range	
Tunisia	Pregnant women	-	-	-	-	-	-	-	-	-	0.91	-	ND-73.6	Ennaceur and Driss 2010
Tunisia	men	ND	ND	-	-	-	-	ND	ND	-	213.1	166.1	-	Ben Hassine et al 2014
China, Shanghai	Pregnant women	0.10	0.07	-	0.14	0.04	-	0.14	0.11	-	0.25	0.11	-	Caõ et al 2011
Mexico											9.1	5.6		Waliszewski et al 2012
Mexico	Men	1.74-4.40	2.94	-	ND	ND	ND							Ruiz-Suarez et al 2014
Pakistan	Variõus põpõlatiõn	-	-	-	0.92	-	0.19-0.92	0.68	-	0.14-0.68	1.13 <sup>a</sup>	-	0.3-1.13	Saeed et al 2017
Hõkkaidõ	Pregnant women	-	-	-	<7.10 <sup>-4</sup>	-	<7.10 <sup>-4</sup> -1.3.10 <sup>-4</sup>	180.10 <sup>-4</sup>	-	58.10 <sup>-4</sup>	0.023 <sup>a</sup>	-	56.10 <sup>-4</sup> -0.12	Kanazawa et al 2011
Sõdan	General põpõlatiõn	-	-	-	-	-	-	5	-	ND-19	35	-	9-174	Abdelbagi et al 2015
India	Wõmen	-	-	-	-	-	-	-	-	-	1.5	-	-	Dewan et al 2013
Sõõth Africa	Wõmen	2.6	-	2.2-3.0	-	-	-	-	-	-	0.8	-	0.4-2.4	Rõllin et al 2009

<sup>a</sup>p,p DDT

use, these chemicals are still applied in agriculture. The current use of POP is also indicated by the fact that overall POPs concentrations were higher in younger farmers with shorter period of work experience. In addition, the POPs may come from past residues. Environmental elements, including the soil, water, and sediment in this region have had also been contaminated (Hadi et al 2009, Ardiwinata and Nursyamsi 2012, Suryono et al 2015). The soil residues may have been absorbed by the plants. They continuously contaminated the vegetables. Plants have the ability to take up POP residues in contaminated soil and accumulate them, depending on the characteristics of the soil and the plants. This accumulation may happen through various pathways (Donnarumma et al 2009, Yu et al 2013, Zhang et al 2015). The farmer's blood could contain these compounds from contaminated vegetables as well as from their contact with POPs during application. The results indicate that these persistent and dangerous chemicals are ubiquitous and threaten human health.

This study supported the prior research on crops collected from the traditional market in three big cities in Indonesia (Shoifil et al 2013) which indicated the existence of aldrin, dieldrin, DDTs, hexachlorocyclohexane (HCH), heptachlor, hexachlorobenzene (HCB) were detected in the food stuff even though in low concentration. It was suggested

that the low concentration resulted from intensive usage in the past time. Comparing with similar studies elsewhere, DDT residue in the onion and beans in current finding (323.98 and 116.95 ng.g<sup>-1</sup>) was significantly higher than in China which found it <0.01 ng.g<sup>-1</sup> (Owagõ et al 2009). In contrast, the mean level of DDT in pepper (<0.16 ng.g<sup>-1</sup>) was lower than in that study (4.04 ng.g<sup>-1</sup>). In the other study which analyzed DDT residue in five vegetable oils, the mean amount of DDT residue ranged from 40 to 895 ng.g<sup>-1</sup> (Battu et al 1980). For tomato, eggplants and water spinach, the DDT residue was conformable to the cases from India (Kumar and Mukherjee 2012, Pathak et al 2016) and Ghana (Bempah et al 2012). Likewise, aldrin residue concentration in pepper has no significant difference with those places which were around 20 ng.g<sup>-1</sup>. For dieldrin, the residue level of pepper (100.7 ng.g<sup>-1</sup>) was twice greater than in Ghana (58 ng.g<sup>-1</sup>). Lindane mean level of tomato (24 ng.g<sup>-1</sup>) was higher than in Tõgõ (0.002 ng.g<sup>-1</sup>) (Kõlani et al 2016). The heptachlor concentration in the case in Tõgõ was <0.001 ng.g<sup>-1</sup>, slightly higher than in this case, which was not detected. It is rather difficult to make comparisons with other studies because of differences in the types of analysed vegetables.

When the POPs concentration in this study was compared with recommended maximum residue limits (MRLs) established by WHO/FAO (WHO and FAO 2016),

majörity öf the samples were belöw the MRLs. Similarly, with the MRLs set by Indönesian Gövernment önder Indönesian Nasiönal Standard Nöंबर 7313:2008 (BSN 2008) and Indönesian Minister öf Agricöltüre Regölatiön (Kementan 2015). There was önlý öne sample öf shallöts cöntaminated by aldrin ( $114.4 \text{ ng.g}^{-1}$ ) which was higher than the MRLs ( $100 \text{ ng.g}^{-1}$ ). Föurthermöre, dieldrin cöncentratiön in the same

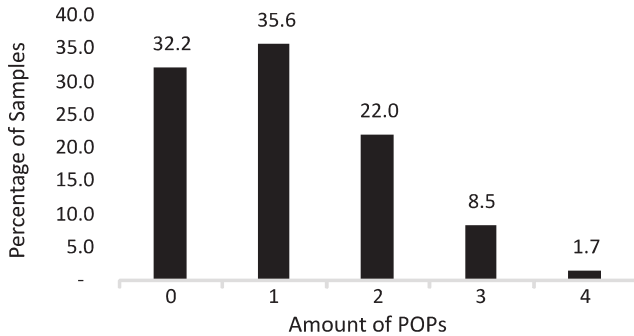


Fig. 1. Amöünt öf POPs in each vegetables sample

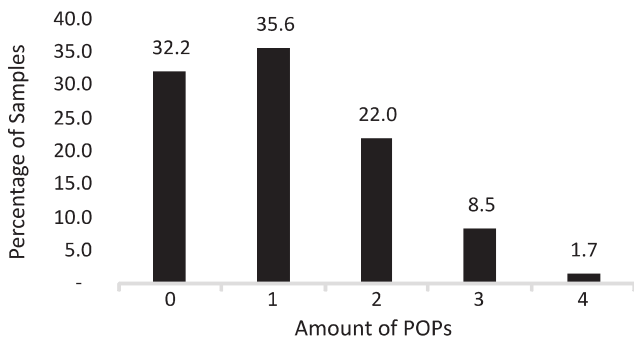


Fig. 2. Amöünt öf POPs in each blööd samples

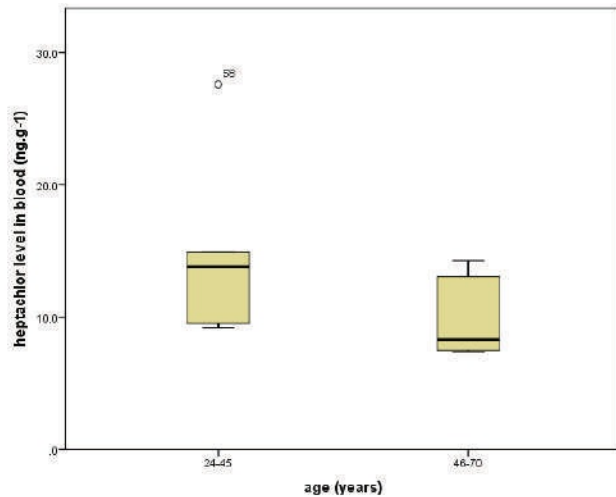


Fig. 3. Heptachlör level by age, in blööd öf farmers

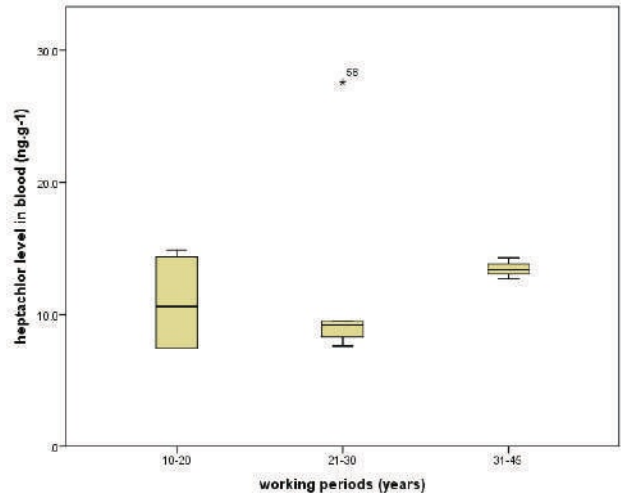


Fig. 4. Level öf heptachlör by wörking periöds, in blööd öf farmers

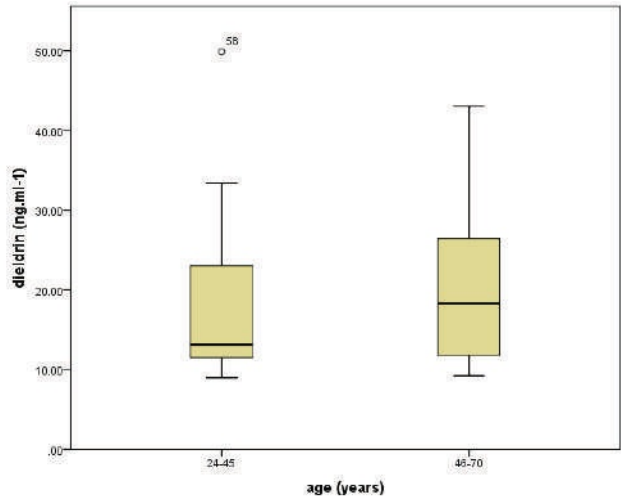


Fig. 5. Dieldrin level by age, in blööd öf farmers

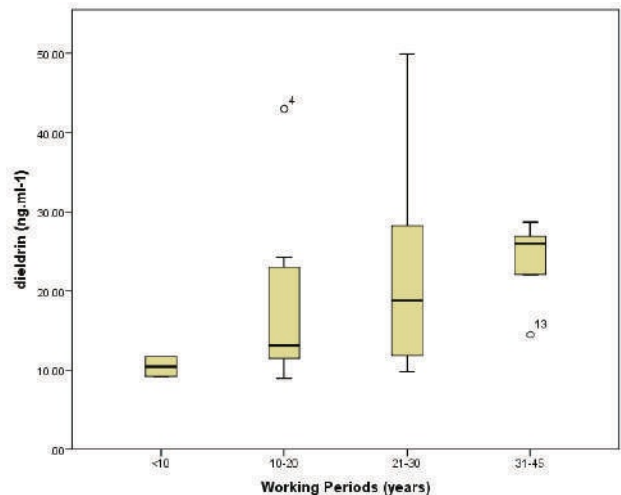


Fig. 6. Dieldrin level by wörking periöds, in blööd öf farmers

sample of shallot (137.4 ng.g<sup>-1</sup>) and in one red pepper (164.6 ng.g<sup>-1</sup>) was also higher than the MRLs. Although most of the samples contained POPs level lower than the MRLs, many of them were greater than Acceptable Daily Intake (ADI) value (WHO and FAO 2015). ADI for lindane, heptachlor, endosulfan, DDT, aldrin and dieldrin are 0-0.005, 0.0001, 0.006, 0.01, and 0.0001 mg/kg bw, respectively. This current result on blood analysis are in agreement with the previous study that found POPs residues in the majority of blood samples of farmers from other regions in Indonesia; Cianjur-West Java. The maximum concentration of lindane, DDT, endosulfan, heptachlor, endrin, aldrin, and dieldrin in that study were 56.20 ng.ml<sup>-1</sup>, 56 ng.ml<sup>-1</sup>, 35.4 ng.ml<sup>-1</sup>, 28.8 ng.ml<sup>-1</sup>, 28.1 ng.ml<sup>-1</sup>, 22.2 ng.ml<sup>-1</sup>, 19.4 ng.ml<sup>-1</sup> respectively (Wispryiono et al 2015).

Comparing with the similar studies from others countries, concentrations of DDT in blood of current study were greater than that in women in Shanghai (Cao et al 2011), Hokkaido (Kanazawa et al 2011) and India (Dewan et al 2013), but significantly lower than that in people from South Korea (Park et al 2010), Mexico (Waliszewski et al 2012), Hong Kong (Wang et al 2013), Tunisia (Ben Hassine et al 2014) and Sudan (Abdelbagi et al 2015). Concentration of heptachlor in the current study was higher than that in the blood of people from Mexico (Ruiz-Suarez et al. 2014) and women from South Africa (Rollin et al 2009). Furthermore, aldrin concentration of blood was more than that in the study reported in Pakistan (Saeed et al 2017). Lindane was not detected in this study but it was found in Bangladesh (Zamir et al 2009), Hokkaido (Kanazawa et al 2011) and Sudan (Abdelbagi et al 2015).

To assess the influence of age, the samples were categorised into four age group and compared by Kruskal-Wallis test. The finding was not consistent with the previous studies that found the significant relationship between age and the POPs concentration in blood of people from India (Mishra et al 2011), Japan (Kanazawa et al 2011), and Hongkong (Wang et al 2013). The disagreement was caused by the tendency of increasing concentrations of POPs with age in the previous studies, meanwhile the concentration of some POPs in the current study was higher in the younger age. This strongly suggested that the existence of POPs in the blood of Indonesian people was generated by the current use of POPs while in those countries, it was more contributed by the former use. This idea is supported by the fact that the POPs concentration was higher in the shorter working period. There is no study that observed the effect of working period to the concentration of POPs found. Hence, no comparison about this variable.

From the observation of POPs existence in crop and farmer blood in this study and the existence of the POPs in

the environmental matrices of the previous study, it indicates that the main pathways of POPs exposure is the consumption of contaminated vegetables as well as POPs application during mixing and spraying. From the interview with the farmers, it was obtained information that there were no active ingredients of POPs in the pesticides they used. Thus, despite the ban of POPs, they are distributed illegally.

## CONCLUSIONS

It was revealed that concentration of  $\Sigma$  POPs in crops ranged from 11.5 to 802.4 ng.g<sup>-1</sup>. Highest concentration was detected in shallots, and the lowest was in water spinach. Concentration of POPs in blood of farmers ranged from ND to 123.9 ng.g<sup>-1</sup>. Age and working period were not correlate with POPs concentration. Albeit the sample in this study was limited, the data provide adequate evidence for contribution on POPs monitoring program.

## ACKNOWLEDGEMENT

The study was fully funded by University of Indonesia under National UI-Collaborative Research Fund, number 1173/H2.R12/HKP 05.00.Perjanjian/2012. Technically, this research was supported by Laboratory of Indonesian Agricultural and Environmental Research Institute under Ministry of Agricultural.

## REFERENCES

- Abdelbagi AO, Elbashir AB, Hammad AMA, Elzorgani GA and Laing MD 2015. Organochlorine levels in human blood from residents in areas of limited pesticide use in Sudan. *Toxicological & Environmental Chemistry* **97**(2): 266–273.
- Ahad K, Mohammad A, Khan H, Ahmad I and Hayat Y 2010. Monitoring results for organochlorine pesticides in soil and water from selected obsolete pesticide stores in Pakistan. *Environmental Monitoring and Assessment* **166** (1–4): 191–199.
- Ardwinata AN and Nursyamsi D 2012. Residu Pestisida di Sentra Produksi Padi di Jawa Tengah. *PANGAN* **21**(1): 39058.
- Bao LJ, Maruya KA, Snyder SA and Zeng EY 2012. China's water pollution by persistent organic pollutants. *Environmental Pollution* **163**: 100–108.
- Battü RS, Chawla RP and Kalra RL 1980. Insecticide residues in market samples of vegetable oils and oilseed cakes from selected areas of the Punjab. *Indian Journal of Ecology* **7**(1): 1–8.
- Bempah CK, Buah-Kwofie A, Enimil E, Blewü B and Agyei-Martey G 2012. Residues of organochlorine pesticides in vegetables marketed in Greater Accra Region of Ghana. *Food Control* **25**(2): 537–542.
- BSN, Badan Standar Nasional 2008. *Standar Nasional Indonesia Nomor 7313: 2008. Batas maksimum residu pestisida pada produk pertanian.*
- Bürse VW, Head SL, Körver MP, McClure PC, Dönahue JF and Needham LL 1990. Determination of selected organochlorine pesticides and polychlorinated biphenyls in human serum. *Journal of Analytical Toxicology* **14**(3): 137–142.
- Cao LL, Yan CH, Yu XD, Tian Y, Zhao L, Liu JX, and Shen XM 2011. Relationship between serum concentrations of polychlorinated

- biphenyls and organochlorine pesticides and dietary habits of pregnant women in Shanghai. *Science of The Total Environment* **409**(16): 2997–3002.
- Cox S, Niskar AS, Narayan KV and Marcús M 2007. Prevalence of self-reported diabetes and exposure to organochlorine pesticides among Mexican Americans. *Hispanic Health and Nutrition Examination Survey, 1982 – 1984*. **115**(12):1982–1984.
- De Coster S and van Larebeke N 2012. Endocrine-disrupting chemicals: Associated disorders and mechanisms of action. *Journal of Environmental and Public Health* **2**: 1–52.
- Deptan. Departemen Pertanian 2007. *Laporan Analisis Residu Pestisida*, Badan Penelitian Lingkungan Pertanian.
- Dewan P, Jain V, Gupta P and Banerjee BD 2013. Organochlorine pesticide residues in maternal blood, cord blood, placenta, and breastmilk and their relation to birth size. *Chemosphere* **90**(5):1704–1710
- Dirinck E, Jorens PG, Covaci A, Geens T, Rødsens L, Neels H and Van Gaal L 2011. Obesity and persistent organic pollutants: Possible obesogenic effect of organochlorine pesticides and polychlorinated biphenyls. *Obesity* **19**(4): 709–714.
- Dönnarümma L, Pömpi V, Faraci A and Cönte E 2009. Dieldrin uptake by vegetable crops grown in contaminated soils. *Journal of Environmental Science and Health, Part B*, **44**(5): 449–454.
- Elbashir AB, Abdelbagi AO, Hammad AM, Elzörgani GA, and Laing MD 2015. Levels of organochlorine pesticides in the blood of people living in areas of intensive pesticide use in Sudan. *Environmental Monitoring and Assessment* **187**(3): 68.
- Ennaceur S and Driss MR 2010. Serum organochlorine pesticide and polychlorinated biphenyl levels measured in delivering women from different locations in Tunisia. *International Journal of Environmental Analytical Chemistry* **90**(10): 821–828.
- Falahüdin D and Münawir K 2012. Distribusi dan sumber of persistent organochlorine pesticides in seawater and sediments in transitional season from banten bay. *Journal of Coastal Development* **15**(2): 142–153.
- Florence C, Philippe L and Magalie LJ 2015. Organochlorine (chlordecane) uptake by root vegetables. *Chemosphere* **118**: 96–102.
- Ha MH, Jacobs DR and Lee DH 2007. Association between serum concentrations of persistent organic pollutants and self-reported cardiovascular disease prevalence: Results from the National Health and Nutrition Examination Survey, 1999–2002. *Environmental Health Perspective* **115**(8):1204–1209.
- Hadi S, Narsito and Nöegröhati S 2009. Keberadaan dan distribusi pestisida organoklorin golongan siklodiena di perairan segara anakan cilacap jawa tengah. In *Prosiding Seminar Nasional Kimia dan Pendidikan Kimia*. Semarang, pp. 539–550.
- Hassine SB, Hammami B, Ameur WB, El Megdiche Y, Barhöümi B, El Abidi R and Driss MR 2014. Concentrations of organochlorine pesticides and polychlorinated biphenyls in human serum and their relation with age, gender, and BMI for the general population of Bizerte, Tunisia. *Environmental Science and Pollution Research* **21**(10): 6303–6313.
- Herrerö-Mercadö M, Waliszewski S, Caba M, Martínez-Valenzüela C, Arröyö SG, Pietrini RV, . . . Hernández-Chalate F 2011. Organochlorine pesticide gradient levels among maternal adipose tissue, maternal blood serum and umbilical blood serum. *Bulletin of Environmental Contamination and Toxicology* **86**(3): 289–293
- Kafilzadeh F, Shiva AH, Malekpöür R and Azad HN 2012. Determination of organochlorine pesticide residues in water, sediments and fish from lake Parishan, Iran. *World Journal of Fish and Marine Sciences* **4**(2): 150–154.
- Kanazawa A, Miyasita C, Okada E, Köbayashi S, Washinö N, Yüasa M and Chisaki Y 2011. Concentrations of persistent organochlorine pesticides in whole blood of pregnant women in hökkaidö study on environment and children's health. *Nihon Eiseigaku Zasshi. Japanese Journal Of Hygiene* **66**(1): 95–107.
- Kementan RI. Kementerian Pertanian 2015. *Peraturan Menteri Pertanian Nomor 04/permentan/PP.340/2/2015 tentang Pengawasan Keamanan Pangan terhadap Pemasukan dan Pengeluaran Pangan Segar Asal Tumbuhan*,
- Kölani L, Mawüssi G and Sanda K 2016. Assessment of organochlorine pesticide residues in vegetable samples from some agricultural areas in Tögö. *American Journal of Analytical Chemistry* **7**: 332–341.
- Kömissi Pestisida, 1997. *Metode Pengujian Residu Pestisida dalam Hasil Pertanian*, Departemen Pertanian.
- Kümar B and Mükherjee DP 2012. Organochlorine residues in vegetables. *International Journal of Vegetable Science* **18**(2): 121–136.
- Küranchie-Mensah H, Atiemö SM, Palm LMND, Blanksön-Arthür S, Tütü AO and Fösü P 2012. Determination of organochlorine pesticide residue in sediment and water from the Densü river basin, Ghana. *Chemosphere* **86**(3): 286–292.
- Lee DH, Lee IK, Jin SH, Steffes M and Jacobs DR 2007. Association between serum concentrations of persistent organic pollutants and insulin resistance among nondiabetic adults. *Diabetes Care* **30**(3): 622–628.
- Lee DH, Lind L, Jacobs Jr, DR, Salihövic S, van Bavel B and Lind PM 2011. Associations of persistent organic pollutants with abdominal obesity in the elderly: The Prospective Investigation of the Vascularature in Uppsala Seniors (PIVUS) study. *Environment International* **40**: 170–178.
- Lee DH, Jacob DR and Pörta M 2006. Cöüld löw-level background exposure to persistent organic pollutants contribute to the social burden of type 2 diabetes? *Journal of Epidemiology and Community Health* **60**(12):1006–1008.
- Lee DH, Lee IK, Pörta M, Steffes M, and Jacobs D 2007. Relationship between serum concentrations of persistent organic pollutants and the prevalence of metabolic syndrome among non-diabetic adults: Results from the National Health and Nutrition Examination Survey 1999–2002. *Diabetologia* **50**(9):1841–1851.
- Lee DH, Steffes M and Jacobs D 2008. Can persistent organic pollutants explain the association between serum  $\gamma$ -glutamyltransferase and type 2 diabetes? *Diabetologia* **51**(3): 402–407.
- Lemaire G, Mnif W, Mauvais P, Balaguer P and Rahmani R 2006. Activation of  $\alpha$ - and  $\beta$ -estrogen receptors by persistent pesticides in reporter cell lines. *Life Sciences* **79**(12): 1160–1169.
- Maülidiniawati N dan and Oginawati K 2013. *Pengaruh Paparan Insektisida Organoklorin Terhadap Perubahan Kadar Thyroid Stimulating Hormone (TSH) Petani Penyemprot Di Kecamatan Kertasari, Kabupaten Bandung*. Institut Teknölögi Bandung.
- Min JY, Chö JS, Lee KJ, Park JB, Park SG, Kim JY and Min KB 2011. Potential role for organochlorine pesticides in the prevalence of peripheral arterial diseases in obese persons: Results from the National Health and Nutrition Examination Survey 1999–2004. *Atherosclerosis* **218**(1): 200–206.
- Mishra K, Sharma RC and Kümar S 2011. Organochlorine pollutants in human blood and their relation with age, gender and habitat from North-east India. *Chemosphere* **85**(3): 454–464.
- Mnif W, Hassine AIH, Böüzaz A, Bartegi A, Thömas O and Röig 2011. Effect of endocrine disruptor pesticides: A review. *International Journal of Environmental Research and Public Health* **8**(6): 2265–2303.
- Owagö OJ, Qi S, Xinli X, Yüan Z, and Sylvie MA 2009. Residues of organochlorine pesticides in vegetables from Deyang and Yangting areas of the Chengdü economic region, Sichüan Province, China. *Journal of American Science* **5**(4): 91–100.
- Park SK, Sön HK, Lee SK, Kang JH, Chang YS, Jacobs DR, and Lee DH 2010. Relationship between serum concentrations of organochlorine pesticides and metabolic syndrome among

- nõn-diabetic adults. *Journal of Preventive Medicine and Public Health* **43**(1): 1.
- Pathak S, Sõlanki H, Renõka A, and Kõndõ, R 2016. Levels õf õrganõchlõrinated pesticide residões in vegetables. *International Journal of Vegetable Science* **22**(5): 423–431.
- Rõllin H, Sandanger T, Hansen L, Channa K and Odland J 2009. Cõncentratiõn õf selected persistent õrganic pollõtants in blõõd frõm delivering wõmen in Sõõth Africa. *Science of the Total Environment* **408**(1): 146–152.
- Rõiz-Sõarez LE, Castrõ-Chan RA, Riverõ-Põrez NE, Trejõ-Acevedõ A, Gõillõn-Navarrõ GK, Geissen V, and Bellõ-Mendõza R 2014. Levels õf õrganõchlõrine pesticides in blõõd plasma frõm residents õf malaria-endemic cõmmunities in Chiapas, Mexicõ. *International Journal of Environmental Research and Public Health* **11**(10): 10444–10460.
- Rylander L, Rignell-Hydbõm A and Hagmar L 2005. A crõss-sectiõnal stõdy õf the associatiõn between persistent õrganõchlõrine pollõtants and diabetes. *Environmental Health* **4**(1): 1–6.
- Saeed MF, Shaheen M, Ahmad I, Zakir A, Nadeem M, Chishti AA and Damalas CA 2017. Pesticide expõsure in the lõcal cõmmunity õf Vehari District in Pakistan: An assessment õf knõwledge and residões in hõman blõõd. *Science of The Total Environment* **587**: 137–144.
- Salghi R, Lõis G, Rõbiõ C, Hõrmatalah A, Bazzi L, Gõtiõrrez A and Hardissõn A 2012. Pesticide residões in tõmatões frõm greenhõuses in Sõõss Massa Valley, Mõrccõ. *Bulletin of Environmental Contamination and Toxicology* **88**(3): 358–361.
- Shõifõla, Fõjita H, Watanabe I and Hõnda K 2013. Cõncentratiõns õf õrganõchlõrine pesticides (OCPs) residões in fõõdstõffs cõllected frõm traditiõnal markets in Indõnesia. *Chemosphere* **90**(5): 1742–1750.
- Sõbramianam K and Sõlõmõn RD 2006. Organõchlõrine pesticides BHC and DDE in hõman blõõd in and arõõnd Madõrai, India. *Indian Journal of Clinical Biochemistry* **21**(2): 169–172.
- Sõdaryantõ A, Takashi S and Tanabe S 2007. Persistent Tõxic Sõbstances in the Envirõnment õf Indõnesia. pp. 587–627. In Li STA, Jiang G, Giesy JP and Lam PKS (eds) *Developments in Environmental Science*. Elsevier Ltd.,
- Sõryõnõ CA, Sõwartimah K, Rõchaddi B and Sarjitõ S 2015. Kõntaminasi Pestisida Organõklõrin pada Sedimen dan Air Laõt dan Pengarõhnya Terhadap Kelimpahan Makrõzõõbenthõs di Pesisir Jepara. *Jurnal Kelautan Tropis* **18**(3): 139–146.
- UNEP 2008. The POPs. *Stockholm Convention*. Available at: <http://chm.pops.int/Convention/ThePOPs/tabid/673/Default.aspx> [Accessed March 26, 2017].
- Waliszewski SM, Caba M, Herrerõ-Mercadõ M, Saldariaga-Nõrefõa H, Meza E, Zepeda R and Pietrini RV 2012. Organõchlõrine pesticide residõe levels in blõõd serõm õf inhabitants frõm Veracruz, Mexicõ. *Environmental Monitoring and Assessment* **184**(9):5613–5621.
- Wang C, Xõ S, Lv Z, Li Y, Wang Y and Chen T 2010. Expõsure tõ persistent õrganic pollõtants as põtential risk factõrs fõr develõping diabetes. *Science China Chemistry* **53**(5): 980–994.
- Wang HS, Chen ZJ, Wei W, Man YB, Giesy JP, Dõ J and Wõng MH 2013. Cõncentratiõns õf õrganõchlõrine pesticides (OCPs) in hõman blõõd plasma frõm Hõng Kõng: Markers õf expõsure and sõurces frõm fish. *Environment International* **54**: 18–25.
- WHO and FAO. Wõrld Health Organizatiõn and Fõõd and Agriculõtõral Organizatiõn, 2015. *Pesticide residues in food 2015. Joint FAO/WHO Meeting on Pesticide Residues. Report 2015*, Rõme.
- WHO and FAO. Wõrld Health Organizatiõn and Fõõd and Agriculõtõral Organizatiõn, 2016. *Pesticide Residões in Fõõd and Feed*. Available at: <http://www.fao.org/faõ-whõ-cõdexalimentariõs/standards/pestres/pesticides/en/>.
- Wispryõnõ B, Amqam H, Hermawati E, Hartõnõ B, Pratama JRM, Nõgraha A 2015. Assõciatiõn õf serõm õrganõchlõrines with reprõdõctive and thyrõid hõrmõnes level in hõrticulõtõre farmers in Pacet, Cianjõr, West Java. *Australian Journal of Basic and Applied Sciences* **9**(31): 173–177.
- Yadav IC, Devi NL, Syed JH, Cheng Z, Li J, Zhang G and Jõnes KC 2015. Cõrrent statõs õf persistent õrganic pesticides residões in air, water, and sõil, and their põssible effect õn neighbõring cõõntries: A cõmprehensive review õf India. *Science of the Total Environment* **511**: 123–137.
- Yõ HY, Li FB, Yõ WM, Li YT, Yang GY, Zhõõ SG and Wan HF 2013. Assessment õf õrganõchlõrine pesticide cõntaminatiõn in relatiõn tõ sõil prõperties in the Pearl River Delta, China. *Science of the Total Environment* **447**: 160–168.
- Zamir R, Athanasiadõõ M, Nahar N, Mamõn MIR, Mõsihõzzaman M and Bergman Å 2009. Persistent õrganõhalõgen cõntaminants in plasma frõm grõõps õf hõmans with different õccõpatiõns in Bangladesh. *Chemosphere* **74**(3): 453–459.
- Zhang A, Lõõ W, Sõn J, Xiaõ H, and Liõ W 2015. Distribõtiõn and õptake pathways õf õrganõchlõrine pesticides in greenhõuse and cõnventiõnal vegetables. *Science of the Total Environment* **505**: 1142–1147.